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EFFECT OF LOW‐TEMPERATURE PYROLYSIS CONDITIONS ON BIOCHAR FOR AGRICULTURAL USE

J. W. Gaskin, C. Steiner, K. Harris, K. C. Das, B. Bibens

ABSTRACT. *The removal of crop residues for bio‐energy production reduces the formation of soil organic carbon (SOC) and therefore can have negative impacts on soil fertility. Pyrolysis (thermoconversion of biomass under anaerobic conditions) generates liquid or gaseous fuels and a char (biochar) recalcitrant against decomposition. Biochar can be used to increase SOC and cycle nutrients back into agricultural fields. In this case, crop residues can be used as a potential energy source as well as to sequester carbon (C) and improve soil quality. To evaluate the agronomic potential of biochar, we analyzed biochar produced from poultry litter, peanut hulls, and pine chips produced at 400°C and 500°C with or without steam activation.* The C content of the biochar ranged from 40% in the poultry litter (PL) biochar to 78% in the pine chip (PC) biochar. The
total and Mehlich I extractable nutrient concentrations in the biochar were strongly influenced by f *total and Mehlich I extractable nutrient concentrations in the biochar were strongly influenced by feedstock. Feedstock A large proportion of N was conserved in the biochar, ranging from 27.4% in the PL biochar to 89.6% in the PC biochar. The amount of N conserved was inversely proportional to the feedstock N concentration. The cation exchange capacity was significantly higher in biochar produced at lower temperature. The results indicate that, depending on feedstock, some biochars have potential to serve as nutrient sources as well as sequester C.*

Keywords. Agricultural residues, Biochar, Bioenergy, Black carbon, Carbon sequestration, Charcoal, Plant nutrition, Pyrolysis, Soil fertility, Soil organic carbon.

yrolysis of crop residues to produce renewable ener‐ gy is one option to reduce the use of fossil fuels. Py‐ rolysis generates biochar, oil, and gas products that can all be used as fuels (Ioannidou and Zabaniotou, Pyrolysis of crop residues to produce renewable ener-
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can all be used as fuels (Ioannidou and Zabaniotou,
2007). cost sorbent (Ioannidou and Zabaniotou, 2007) or as a soil amendment to improve soil fertility and sequester carbon (Lehmann et al., 2006; Steiner, 2007). Removal of crop resi‐ dues for energy production can have deleterious effects on soil organic carbon (SOC) and consequently on soil fertility (Lal, 2004). Pyrolysis of crop residues with C returned to the soil in the form of biochar may help maintain or increase stable SOC pools and cycle nutrients back into agricultural fields. Pyrolysis with biochar C sequestration may offer an option to reduce the conflict between cultivating crops for different purposes, e.g., energy vs. C sequestration or food.

There are several lines of evidence that charcoal plays an important role in soil fertility. Charcoal has been identified as an important soil constituent in fertile Chernozems (Schmitdt et al., 1999) and in anthropogenic enriched dark soil (Terra Preta) found throughout the lowland portion of the Amazon Basin (Glaser et al., 2000). Research on tropical soils indicates that charcoal amendments can increase and sustain soil fertility (Steiner et al., 2007). The beneficial effects appear to be related to alterations in soil physical, chemical, and biological properties, such as reduced acidity (Topoliantz et al., 2005), increased cation exchange capacity (CEC) (Cheng et al., 2008; Liang et al., 2006), enhanced ni‐ trogen (N) retention (Lehmann et al., 2003; Steiner et al., 2008b), increased microbiological activity (Steiner et al., 2008a), and increased mycorrhizal associations (Warnock et al., 2007). Research on the effect of wildfire charcoal in for‐ est ecosystems indicates that it stimulates microbial activity (Pietikäinen et al., 2000) and influences nitrogen cycling (Berglund et al., 2004; DeLuca et al., 2006; Wardle et al., 1998). Research also indicates that charcoal is recalcitrant (Seiler and Crutzen, 1980), and it may persist for hundreds or thousands of years.

Charcoals produced from wildfire or traditional charcoal production may have different chemical and physical charac‐ teristics from pyrolytic biochars created under specific conditions for energy production. Both feedstock and pyrolysis conditions such as temperature and carrier gas affect the chemical and physical characteristics of biochar (Antal and Grønli, 2003; Bansal et al., 1988; Benaddi et al., 2000; Guo and Rockstraw, 2007a; Strelko et al., 2002). Most of the liter-
ature discusses high-temperature biochars that are produced
at greater than 500°C or activated carbon typically produced ature discusses high‐temperature biochars that are produced at greater than 500°C or activated carbon typically produced
at 800°C. As pyrolysis temperatures increase, volatile compounds in the biochar matrix are lost, surface area and ash in‐ crease, but surface functional groups that can provide exchange capacity decrease (Guo and Rockstraw, 2007a).

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Pyrolysis of nutrient‐rich feedstock is likely to produce nutrient‐rich biochar, but nutrient conservation and availabil‐ ity in biochars is not well understood. Nutrients susceptible to volatilization such as N are almost completely lost after a burn (Giardina et al., 2000). Whether elements are retained during pyrolysis, the availability of nutrients for plants, and the effect of pyrolysis conditions on these characteristics are unclear. For biochar to be used in agriculture, a better under‐ standing of its properties and how it affects soil fertility is needed. Therefore, our objectives were to determine the effect of feedstock, temperature, and carrier gas on key characteristics of biochar for agricultural use. Specifically, we wished to compare characteristics critical for agricultural use including pH, CEC, total nutrient concentrations, and poten‐ tially available nutrient concentrations in biochars from three feedstocks under two temperature regimes using two carrier gases with and without secondary steam activation.

MATERIALS AND METHODS

BIOCHAR PRODUCTION

We selected three common feedstocks to represent a range of physical properties and mineral content: raw poultry litter from broiler houses (*Gallus domesticus*, PL), pelletized pea‐ nut hulls (*Arachis hypogaea*, PN), and raw pine chips (*Pinus taeda*, PC). Biochars were produced in a batch pyrolysis unit nut hulls (*Arachis hypogaea*, PN), and raw pine chips (*Pinus taeda*, PC). Biochars were produced in a batch pyrolysis unit at two peak temperatures (400 $^{\circ}$ C and 500 $^{\circ}$ C) with either steam or nitrogen (N₂) as a carrier gas. The biochars produced with N₂ as a carrier gas were produced with or without steam activation at the original pyrolysis temperatures (400[°]C and with N_2 as a carrier gas were produced with or without steam activation at the original pyrolysis temperatures (400 $^{\circ}$ C and ⁵⁰⁰°C). Each of the production combinations (three feed‐ stocks, three pyrolysis types, two temperatures $= 18$) was replicated three times. The conversion efficiency was calculated as the percentage of the feedstock input (dry weight, DW) and biochar output (biochar DW / feedstock DW).

CHEMICAL ANALYSES

Biochars were ground in a ball mill to pass a 300 µm sieve before nutrient analysis. Feedstock and the biochars were analyzed for total C, N, and sulfur (S) by dry combustion (CNS‐2000, Leco Corp., St. Joseph, Mich.). Total minerals were extracted using a closed‐vessel microwave digestion with $HNO₃$ (USEPA method 3050; USEPA, 1994). A Mehlich I extraction (0.05 M HCl + 0.0125 M H₂SO₄) (Mehlich, 1953) was also used on biochar samples as an index of poten‐ tially plant‐available nutrients. Aluminum, Cu, Ca, Fe, Mg, Mn, P, K, Na, and Zn were measured by inductively coupled plasma spectrometry (ICP, Thermo Jarrell‐Ash model 61E, Thermo Fisher Scientific, Waltham, Mass.).

Biochar pH was measured in deionized water using a 1 to 5 wt/wt ratio. Samples were thoroughly mixed and allowed to equilibrate for 1 h. The pH was measured with a digital pH meter (AR15, Thermo Fisher Scientific, Waltham, Mass.).

Cation exchange capacity of the biochar was measured by a modified ammonium‐acetate compulsory displacement (Sumner and Miller, 1996). Samples were leached with de‐ ionized water five times before starting the CEC extraction to reduce interference from soluble salts. Twenty mL of de‐ ionized water was added to a 1 g sample of biochar in a dis‐ posable nalgene 0.45 µm cellulose nitrile filter flask. The flask was placed on an orbital shaker and shaken at 180 rpm for 5 minutes. The sample was vacuum filtered, and the leachate was saved for further analysis. After the fifth wash, 10mL of Na‐acetate (pH 7) was added to the sample, and the mixture shaken for 10 min. This process was repeated three times to ensure that exchange sites were saturated with Na ions. Biochar samples were then washed three times with ethanol to remove excess Na. Sodium ions were displaced with NH_4 -acetate (pH 7) three times and measured by atomic adsorption (PE 4100ZL, Perkin Elmer, Waltham, Mass.).

The reserved leachate from the five washings (CEC proce‐ dure above) was composited and analyzed for dissolved car‐ bon (DC), dissolved inorganic carbon (DIC), ammonium‐ nitrogen (NH₄-N), and nitrate-nitrogen (NO₃-N). Dissolved carbon and DIC was measured by combustion (Shimadzu TOC‐5050A, Shimadzu, Columbia, Md.). Dissolved organic C (DOC) was calculated by difference (DOC = DC - DIC). Nitrate-nitrogen and NH_4-N were analyzed on an autoanalyzer using cadmium reduction and phenate colorimetric methods (EnviroFlow 3000, Perstorp, Toledo, Ohio).

STATISTICAL ANALYSES

Treatment effects were analyzed by general linear model (GLM) univariate analysis of variance (ANOVA). The detection limit was used for results below the detection limit, if other results were above the limit. This allowed a conservative estimate of the elemental concentration of the biochar. If all results were below the detection limit, then no statistical analysis was performed. Significant differences $(p < 0.05)$ between the feedstock and treatments were separated by the Tukey test. Statistical analyses and plots were performed using SPSS 12.0 and SigmaPlot 8.02 (SPSS, Inc., Chicago, Ill.).

RESULTS AND DISCUSSION

Steam pyrolysis of the peanut hull pellets in the batch reactor presented difficulties due to excessive swelling by the peanut hull feedstock that clogged the reactor. Low-temperature steam pyrolysis in a batch reactor may not be appropriate for this feedstock. Analysis of PC and PL biochars revealed no difference in total nutrients, Mehlich I extractable nutrients, CEC, or pH between steam and N_2 as carrier gas; consequently, we report on the results from pyrolysis with the N_2 carrier gas with or without subsequent steam activation.

INFLUENCE OF FEEDSTOCK

The total element concentrations in the feedstock had the strongest influence on the chemical composition of the

Table 1. Total element concentrations in the three agricultural feedstocks used for pyrolysis at 400°C and 500°C.

	Values in $g \text{ kg}^{-1}$										Values in mg kg^{-1}							
Feedstock						Ca	Mg	Al	Fe	Na	Cd	Cr	- Cu	Mn	B	Mo	Ni	Zn
Poultry litter (PL)				326 45.1 19.5 29.5	5.8	28.0	5.66	6.32	3.91	9.27	1.4°	7.3	381	377.0	49.9	3.5		414
Peanut hulls (PN)		552 13.6 0.61		5.06	0.9	1.84	0.79	0.92	0.42	0.04	${<}1$	2.0	36.5	44.0	15.2	15.6		20.2
Pine chips (PC)	571	0.9	0.08	0.59		0.75	0.21	0.03	0.13	0.04	<1	-2.1		.3.8		1.0	2.0	47.8

Table 2a. Means and standard errors for pH, CEC, and total macronutrient concentrations in poultry litter, peanut

	hull, and pine chip biochars. Feedstock and temperature columns indicate significant differences ($p < 0.05$, $n = 3$).													
		Poultry Litter (PL)					Peanut Hulls (PN)			Pine Chips (PC)				
	400 °C		500 °C		400 °C		500 °C		400 °C		500 °C			
	Biochar	SA	Biochar	SA	Biochar	SA	Biochar	SA	Biochar	SA	Biochar	SA	Feedstock	
pH (S.U.)	10.1 ± 0.04	10.1 ± 0.07	9.74 ± 0.05	9.88 ± 0.09	10.5 ± 0.05	10.5 ± 0.10	10.1 ± 0.02	9.96 ± 0.01	7.55 ± 0.09	7.99 ± 0.09	8.30 ± 0.15	8.10 ± 0.60	PC <pl<pn< td=""><td>Temp.</td></pl<pn<>	Temp.
CEC $\pmod{kg^{-1}}$	61.1 ± 0.73	57.4 ±1.4	38.3 ±1.7	37.0 ±1.6	14.2 ± 0.46	11.7 ±2.04	4.63 ± 0.10	4.46 ± 0.13	7.27 ± 0.54	6.00 ± 0.11	5.03 ± 0.85	6.02 ± 3.12	PC <pn<pl< td=""><td>500<400</td></pn<pl<>	500<400
\mathcal{C} $(g \; kg^{-1})$	392 ± 3.8	399 ±7.4	392 ± 8.6	421 ±23	732 ±14	762 ± 3.4	804 ±1.7	806 ±5.8	739 ±17	761 ± 3.6	817 ±1.9	820 ±17	PL <pn, pc<="" td=""><td>400<500</td></pn,>	400<500
N $(g \; kg^{-1})$	34.7 ±0.79	34.7 ±0.77	30.9 ± 0.89	32.3 ±1.6	24.3 ± 0.18	24.0 ± 0.37	24.8 ± 0.89	24.8 ±0.34	2.55 ± 0.40	1.95 ± 0.06	2.23 ± 0.09	2.20 ± 0.12	PC <pn<pl< td=""><td>500<400</td></pn<pl<>	500<400
P $(g \; kg^{-1})$	30.1 ±0.16	32.2 ± 2.3	35.9 ±1.6	34.8 ±2.6	1.83 ± 0.11	1.70 ± 0.12	1.97 ± 0.03	2.06 ± 0.11	0.15 ±0.004	0.14 ±0.004	0.14 ± 0.02	0.20 ± 0.02	PC <pn<pl< td=""><td>400<500</td></pn<pl<>	400<500
K $(g \; kg^{-1})$	51.1 ±1.3	52.6 ±4.9	58.6 ±2.9	54.7 ±1.5	15.2 ± 0.58	14.40 ± 1.40	16.4 ± 0.19	16.5 ± 0.79	1.45 ± 0.06	1.51 ± 0.07	1.45 ± 0.18	2.25 ±0.25	PC <pn<pl< td=""><td>400<500</td></pn<pl<>	400<500
Ca $(g \; kg^{-1})$	42.7 ± 0.30	45.7 ± 3.0	50.4 ±2.2	49.1 ± 3.7	4.62 ± 0.06	4.46 ± 0.29	5.12 ± 0.12	5.21 ± 0.20	1.71 ± 0.11	1.69 ± 0.02	1.85 ± 0.14	2.17 ± 0.04	PC <pn<pl< td=""><td>400<500</td></pn<pl<>	400<500
Mg $(g \; kg^{-1})$	10.7 ±0.23	11.4 ± 0.91	12.9 ± 0.50	12.4 ± 1.0	2.19 ± 0.06	2.17 ± 0.16	2.50 ± 0.05	2.59 ± 0.11	0.60 ± 0.04	0.58 ± 0.03	0.59 ± 0.06	0.76 ± 0.01	PC <pn<pl< td=""><td>400<500</td></pn<pl<>	400<500
S $(g \; kg^{-1})$	13.67 ± 0.39	12.3 ± 0.09	13.93 ±1.1	13.9 ± 0.37	0.56 ± 0.02	0.51 ± 0.03	0.55 ± 0.09	0.37 ± 0.09	0.01 ± 0.04	0.16 ± 0.05	0.06 ± 0.01	0.08 ± 0.04	PC, PN <pl< td=""><td></td></pl<>	

 $\overline{[a]}$ SA = steam activation.

Table 2b. Means and standard errors for total micronutrient and selected element concentrations in poultry litter, peanut

	hull, and pine chip biochars. Feedstock and temperature columns indicate significant differences ($p < 0.05$, $n = 3$). [a]													
		Poultry Litter (PL)				Peanut Hulls (PN)				Pine Chips (PC)			Feedstock	Temp.
		400 °C 500 °C			400 °C		500 °C		400 °C			500 °C		
	Biochar	SA	Biochar	SA	Biochar	SA	Biochar	SA	Biochar	SA	Biochar	SA		
Al	9.87	8.12	13.02	14.25	2.40	2.33	2.73	2.81	0.07	0.05	0.07	0.06		
$(g \; kg^{-1})$	±1.36	±1.59	± 0.36	±1.97	± 0.07	± 0.16	± 0.05	± 0.13	± 0.01	±0.005	± 0.01	± 0.008	PC. PN <pl< td=""><td>400<500</td></pl<>	400<500
Fe	6.06	5.55	8.03	8.89	1.00	0.97	1.15	1.20	0.15	0.04	0.05	0.20		
$(g \; kg^{-1})$	± 0.52	± 0.42	± 0.55	±1.27	± 0.02	± 0.07	± 0.02	± 0.06	± 0.11	±0.007	± 0.01	± 0.07	PC, PN <pl< td=""><td>400<500</td></pl<>	400<500
Na	15.1	15.8	17.2	16.6	0.026	0.028	0.035	0.044	< 0.014	0.053	0.013	0.075		
$(g \; kg^{-1})$	± 0.31	±1.37	±1.02	±1.12	± 0.001	±0.006	±0.004	±0.005	±0.004	±0.032	±0.002	±0.054	PN, PC <pl< td=""><td></td></pl<>	
B	91.5	96.0	100	93.0	32.5	29.9	33.7	34.1	5.69	6.69	4.21	6.94		
$(mg kg-1)$	± 3.16	±8.25	± 0.31	±3.98	±1.57	±2.87	± 0.27	±1.15	± 0.30	± 0.21	± 0.62	± 1.33	PC <pn<pl< td=""><td></td></pn<pl<>	
Cd	2.75	2.65		1.10	< 1.35	< 1.35								
$(mg kg-1)$	± 0.73	± 0.83	<1	± 0.10	± 0.35	± 0.35	<1	<1	<1	<1	<1	<1		
Cr	28.0	28.8	59.4	56.1	3.95	3.00	3.63	3.94		1.23	3.43	17.7		
$(mg kg-1)$	±4.1	±5.2	±3.3	±4.5	± 0.31	± 0.57	± 0.31	± 0.16	<1	± 0.09	± 0.88	±5.9	PN, PC <pl< td=""><td>400<500</td></pl<>	400<500
Cu	805	880	1034	943	16	13	19	19	25	10	9	13		
$(mg kg-1)$	±23	±49	±68	±81	± 0.60	±1.27	± 0.50	±1.84	±7.03	±6.18	±2.34	±5.57	PC, PN <pl< td=""><td>400<500</td></pl<>	400<500
Mn	596	637	725	697	116	116	131	136	274	269	258	350		
$(mg kg-1)$	±5.6	±37	±29	±46	± 2.3	± 8.0	±2.3	±5.7	±9.3	±7.8	±30	±4.0	PN <pc<pl< td=""><td>400<500</td></pc<pl<>	400<500
Mo	17.1	12.1	14.2	13.8	4.78							< 4.11		
$(mg kg-1)$	±5.3	± 0.41	±1.1	±1.2	± 3.6	<1	<1	<1	<1	<1	<1	± 3.11	PC, PN <pl< td=""><td></td></pl<>	
Ni	13.6	19.5	20.3	29.1	2.29	\leq 2	\leq 2	< 10.4	\leq 2	&2	2.91	17.5		
$(mg kg-1)$	± 0.00	±3.7	±1.1	± 8.4	±0.29	±0	±0	± 8.0	±0	±0	± 0.55	±14.7	PC, PN <pl< td=""><td></td></pl<>	
Zn	628	680	752	728	35	31	37	36	15	16	18	20		
$(mg kg-1)$	±12	±41	±28	±50	±2.2	±2.9	± 2.1	± 0.00	±1.1	± 0.7	± 0.6	±2.4	PC, PN <pl< td=""><td>400<500</td></pl<>	400<500

[a] SA = steam activation; < indicates mean contains results below the detection limit; ± 0.00 indicates all results were near instrument detection limit.

biochar. Concentrations of plant nutrients in the feedstocks generally followed the pattern of PC < PN < PL. Feedstock carbon concentrations had the opposite pattern, with PL < PN \leq PC (table 1).

There were significant differences in C concentrations in the biochar, with PL containing less C than the PN or PC biochar (table 2a). The nutrient‐rich poultry litter contains relatively more minerals than the other feedstocks, which

decreases the C content. Nitrogen, P, K, Ca, and Mg concentrations in the biochar were significantly different, with $PC < PN < PL$ (table 2a). The concentration of the micronutrients B, Cu, Fe, Mn, Na, and Zn were significantly higher in PL biochar ($p < 0.05$), but there were no differences detected between the PN and PC biochars except for Mn (table 2b). Concentrations of metals such as Al, Cr, Ni, and Mo were low. The PL biochar contained the highest

Figure 1. Percentages with standard errors of feedstock nutrients conserved in the biochar and percentages of total nutrients that were Mehlich I extractable at two pyrolysis temperatures and in three biochars. Letters above the columns indicate significant difference of nutrients conserved between biochar types (p < 0.05, *n* **= 3). Letters within columns indicate significant difference in the percentage of total nutrients that were Mehlich-I extractable (p < 0.05,** $n = 3$ **).**

concentrations of these metals, as would be expected from the higher feedstock concentrations. Cadmium was below the detection limit in PC biochars and near or at the detection limits in PL and PN biochars (table 2b).

The amount of N conserved ranged from 27.4% in the PL biochar to 89.6% in the PC biochar and was inversely proportional to the feedstock N concentration (fig. 1b and table 1). The higher N losses seen from the PL were likely due to the volatilization of the poultry manure NH_4-N and easily decomposable N-containing organic compounds in the manure, such as uric acid. In contrast, the low concentration of N in the PC feedstock is likely to be incorporated into complex structures that are not easily volatilized.

About 60% of the P in the PL and PC feedstock was retained in the PL and PC biochar, while nearly 100% of the P in the PN feedstock was retained in the PN biochar (fig. 1c).

In general, the PL biochar had a lower proportion of nutrients retained than the PN or the PC biochar (figs. 1c through 1f). This may be due to a higher proportion of some of these elements retained in the aqueous/bio‐oil fraction in PL biochar (K. C. Das, 2007, unpublished data, University of Georgia, Athens, Ga.).

The pattern of Mehlich I extractable concentrations was similar to that of the total nutrient concentrations (tables 3a and 3b). There were significant differences in Mehlich I extractable P, K, Ca, and Mg concentrations, with PC < PN < PL. There were differences by feedstock in the percentage of the total nutrients that were Mehlich I extractable (figs. 1a
through 1f). Only 19% of the PL biochar P was Mehlich I
extractable compared to over 40% in the PN biochar (400°C, through 1f). Only 19% of the PL biochar P was Mehlich I fig. 1c). About 90% of the PL biochar K was Mehlich I extractable compared to over 40% in the PN biochar (400° C, fig. 1c). About 90% of the PL biochar K was Mehlich I extractable compared to only 45% in the PN biochar (400° C,

Table 3a. Means and standard errors of the Mehlich I macronutrient concentrations in poultry litter, peanut hull,

	and pine chip biochars. Feedstock and temperature columns indicate significant differences ($p < 0.05$, $n = 3$). [a]														
		Poultry Litter (PL)					Peanut Hulls (PN)			Pine Chips (PC)		Feedstock	Temp.		
	400 °C		500 °C		400 °C		500 °C		400 °C		500 °C				
	Biochar	SA	Biochar	SA	Biochar	SA	Biochar	SA	Biochar	SA	Biochar	SA			
P	5.58	4.09	5.33	4.66	0.76	0.67	0.57	0.59	0.03	0.034	0.04	0.06			
$(g \; kg^{-1})$	± 0.31	±1.22	± 0.18	± 0.20	± 0.02	± 0.06	± 0.04	± 0.03	± 0.002	± 0.004	± 0.008	± 0.03	PC <pn<pl< td=""><td></td></pn<pl<>		
K	46.2	34.1	38.1	40.0	6.84	6.28	5.91	6.76	0.30	0.38	0.41	0.97			
$(g \; kg^{-1})$	±0.96	± 8.40	± 2.68	± 2.81	± 0.16	± 0.67	± 0.28	± 0.30	± 0.009	± 0.02	± 0.06		± 0.32 PC <pn<pl< td=""><td></td></pn<pl<>		
Ca	3.34	1.95	2.21	1.63	1.68	1.48	1.19	1.22	0.30	0.31	0.43	0.39			
$(g \; kg^{-1})$	±0.84	±0.82	± 0.36	± 0.13	± 0.02	±0.15	± 0.06	± 0.06	± 0.04	± 0.05	± 0.10	±0.16	PC <pn<pl< td=""><td></td></pn<pl<>		
Mg	3.09	2.19	3.03	2.92	0.80	0.62	0.37	0.39	0.05	0.06	0.06	0.08			
$(g \; kg^{-1})$	±0.28	±0.68	± 0.13	± 0.05	± 0.03	± 0.09	± 0.02	± 0.01	± 0.008	± 0.009	± 0.01	± 0.04	PC <pn<pl< td=""><td></td></pn<pl<>		

 $[a]$ SA = steam activation.

Table 3b. Means and standard errors of the Mehlich I micronutrient and selected element concentrations in poultry litter,

	peanut hull, and pine chip biochars. Feedstock and temperature columns indicate significant differences ($p < 0.05$, $n = 3$).[a]													
		Poultry Litter (PL)				Peanut Hulls (PN)				Pine Hips (PC)			Feedstock	Temp.
	400 °C		500 °C		400 °C			500 °C	400 °C		500 °C			
	Biochar	SA	Biochar	SA	Biochar	SA	Biochar	SA	Biochar	SA	Biochar	SA		
Al	0.47	11.3	1.43	1.52	330	585	1129	1360	6.17	6.35	7.53	12.12		
$(g \; kg^{-1})$	± 0.11	±2.24	± 0.31	± 0.24	±24	±213	±32	±51	± 0.59	± 0.53	±1.12	±4.76	PL, PC <pn< td=""><td>400<500</td></pn<>	400<500
Fe	0.66	3.16	0.06	0.19	140	142	197	221	3.72	4.58	14.6	33.3		
$(g \; kg^{-1})$	± 0.21	± 0.08	± 0.007	$n=1$	±2.2	±18	±3.8	±11	± 0.26	± 0.19	±2.89	±8.39	PL, PC <pn< td=""><td>400<500</td></pn<>	400<500
Na	9.57	7.08	6.98	7.24	0.02	0.02	0.02	0.03	0.03	0.03	0.03	0.08		
$(g \; kg^{-1})$	± 0.19	±1.61	± 0.43	± 0.07	± 0.001	±0.002	± 0.58	±0.002	±0.002	± 0.002	±0.005	±0.037	PN, PC <pl< td=""><td></td></pl<>	
\overline{B}	16.7	18.8	18.67	20.4	4.20	4.96	3.97	5.84	0.45	0.41	0.52	1.15		
$(mg kg-1)$	±1.59	± 2.05	± 0.85	±0.59	± 0.12	± 1.08	± 0.22	± 0.66	± 0.04	± 0.07	± 0.07	± 0.49	PC <pn<pl< td=""><td></td></pn<pl<>	
Cr	0.19	0.19	0.14	0.11			0.41	0.52						
$(mg kg^{-1})$	± 0.03	± 0.02	± 0.01	± 0.01	< 0.04	< 0.04	± 0.04	± 0.03	<0.06	< 0.06	< 0.06	< 0.06		
Cu	0.40	0.29	< 0.08	< 0.05	0.67	< 0.59	< 0.04		6.55	2.48	2.70	3.82		
$(mg kg-1)$	± 0.06	± 0.08	± 0.02	±0.005	± 0.04	± 0.32	± 0.001	< 0.04	± 2.18	± 0.68	± 0.89	±3.34	PL, PN <pc< td=""><td></td></pc<>	
Mn	7.69	8.64	6.75	5.17	24.7	21.2	14.4	16.3	22.6	25.2	24.1	36.2		
$(mg kg-1)$	±1.23	±1.28	± 1.03	± 0.43	± 0.70	±2.33	±0.45	± 0.71	± 2.61	±3.79	± 6.67	±12.5	PL <pn<pc< td=""><td></td></pn<pc<>	
Mo	0.87	1.11	1.42	1.94					0.15	0.25	0.11	0.65		
$(mg kg-1)$	± 0.19	± 0.22	± 0.14	± 0.19	< 0.04	< 0.04	< 0.04	< 0.04	±0.002	± 0.06	± 0.03	±0.35		
Ni														
$(mg kg-1)$	< 0.08	<0.08	< 0.08	< 0.08	<0.08	< 0.08	< 0.08	<0.08	<0.2	< 0.2	< 0.2	< 0.2		
Zn	0.06	0.30	0.05		10.51	7.36	5.58	6.30	2.20	2.31	1.36	3.66		
$(mg kg-1)$	± 0.01	± 0.02	± 0.07	< 0.04	±1.22	± 0.59	± 0.26	± 0.45	± 0.17	±0.21	± 0.26	±1.17	PL <pc<pn< td=""><td></td></pc<pn<>	
$T_{\rm eff}$														

 $\lbrack a \rbrack$ SA = steam activation; < indicates mean contains results below the detection limit.

fig. 1d). Manganese and Zn concentrations were significantly lower in the PL biochar than the PC or PN biochars. Copper, Al, and Fe was also lower in the PL biochar compared to the PN biochar. These patterns are the reverse of that seen in the feedstock or the total element concentrations in the biochars. If pyrolysis can reduce P and other metal availability in poultry litter, it may reduce some of the environmental concerns associated with land application of poultry litter. These results should be interpreted with caution. The Mehlich I extraction, which is a weak double acid extraction, may not have been strong enough to remove these acidsoluble cations under the high pH conditions found in the PL biochar.

The Mehlich I extraction was developed for acidic soils in the southeastern U.S. with low CEC or base saturation (Kuo, 1996), and it is the standard extraction used for plant‐ available nutrients and fertilizer recommendations in Alabama, Georgia, Florida, South Carolina, Tennessee, and Virginia. In this study, Mehlich I extractable element concentrations were used as an index to compare the potential for different biomass sources and production techniques to supply plant‐available nutrients. The extraction has not been calibrated for biochar and may not reflect actual plant‐available nutrient concentrations. However, data from a greenhouse trial using pine chip and peanut hull biochar amendment of three different Ultisols (Speir, 2008) and from a field trial with the same biochars (Gaskin et al., 2007) indicate an increase in Mehlich I K and Mg in soils amended with peanut hull biochar. The increased Mehlich I K in the soil was reflected in an increase of these nutrients in corn tissue (*Zea mays*) in the field trial.

The pH and CEC of the biochars were also significantly influenced by feedstock (table 2a). All the biochars were basic, with the highest pH seen in the PN biochar. Tryon (1948) reported increased soil pH with the addition of pine and hardwood charcoal. He attributed the greater pH increase seen in the hardwood charcoal treatment to the higher ash content, in particular to the hydrolysis of salts of Ca, K, and Mg in the presence of water. In this study, PC biochar had both the lowest total concentrations of these cations and the

Table 4. Means and standard errors of dissolved carbon (DC), dissolved inorganic carbon (DIC), dissolved organic carbon (DOC), ammonium‐nitrogen (NH4-N), and nitrate‐nitrogen (NO3-N) in leachate from poultry litter, peanut hull,

	and pine chip biochars. Feedstock and temperature columns indicate significant differences ($p < 0.05$, $n = 3$).														
		Poultry Litter (PL)					Peanut Hulls (PN)			Pine Chips (PC)	Feedstock	Temp.			
	400 °C		500 °C		400 °C		500 °C		400 °C		500 °C				
	Biochar	SA	Biochar	SA	Biochar	SA	Biochar	SA	Biochar	SA	Biochar	SA			
DC	2.20	1.85	0.85	0.75	0.51	0.40	0.52	0.41	0.13	0.13	0.12	0.19			
$(g kg^{-1})$	± 0.31	±0.25	± 0.01	± 0.02	± 0.09	± 0.02	± 0.02	± 0.10	± 0.10	± 0.10	± 0.05	± 0.07	PC <pn<pl< td=""><td>500<400</td></pn<pl<>	500<400	
DIC	0.39	0.44	0.57	0.54	0.32	0.31	0.38	0.37	0.014	0.025	0.034	0.055			
$(g \; kg^{-1})$	± 0.04	± 0.04	± 0.03	± 0.04	± 0.003	± 0.003	± 0.005		± 0.012 PC <pn<pl< td=""><td>400<500</td></pn<pl<>	400<500					
DOC	1.81	1.46	0.28	0.21	0.20	0.10	0.14	0.10	0.12	0.10	0.09	0.10			
$(g \; kg^{-1})$	±0.34	±0.29	± 0.04	± 0.04	± 0.06	± 0.03	± 0.03	± 0.03	± 0.01	± 0.01	± 0.003	± 0.01	PC, PN <pl< td=""><td>500<400</td></pl<>	500<400	
$NH4-N$	8.5	6.69	11.3	3.49	2.86	1.94	2.12	2.28	1.75	7.93	2.41	2.37			
$(mg kg-1)$	±0.39	± 0.69	±6.41	±0.15	± 0.27	± 0.01	± 0.12	± 0.44	± 0.29	±6.16	± 0.16	± 0.08	PN<pl< b=""></pl<>		
$NO3-N$	<0.4	<0.4	<0.4	< 0.4	1.02	1.07	1.27	1.11	< 0.4	< 0.4	< 0.4	< 0.4			
$(mg kg-1)$					± 0.01	± 0.02	± 0.06	± 0.02							

[a] $SA =$ steam activation; < indicates mean contains results below the detection limit.

lowest pH, which would support Tyron's hypothesis; however, the pH of PL biochar was similar to the PN biochar although it contained higher concentrations of total Ca, K, and Mg than PN biochar (table 2).

Higher CEC was associated with higher concentrations of minerals in the feedstock. Mészáros et al. (2007) hypothesized that K, Mg, Na, and P in the biomass may catalyze the formation of oxygen groups on the biochar surface at low pyrolysis temperatures. Oxygen groups such as carboxyls, lactones, and phenols could contribute to the presence of negative surface charges (Boehm, 1994).

Dissolved C concentrations were very low (table 4). Feedstock had a significant effect on DC in the biochar leachate, with PC < PN < PL. Dissolved inorganic C was also affected by feedstock, with PC < PN < PL. The PL biochar had a higher proportion of DOC than PC or PN biochars. The PL feedstock is a combination of wood chip (typically pine) bedding and poultry manure. The manure may contribute to higher DOC leached from the PL biochar.

DOC plays an important role in many soil processes, including serving as an energy source for the microbial community and reacting with other soil solution components (Sposito, 1989). Biochars are known to contain condensed volatile compounds. These compounds are either lost to the gaseous or liquid phase or undergo further reactions to form secondary char at high temperatures (Antal and Grønli, 2003). Garcia‐Perez et al. (2007) identified water‐soluble compounds from pyrolysis of lignin materials to contain mono‐ and oligo‐sugars, formic and aecetic acids, as well as methanol, hydroxyl-acetaaldehyde, and 1-hydroxyl-2monantes, in alternative detailed point in the strength of propanone. Schnitzer et al. (2007) identified numerous organic compounds in the light and heavy fractions of poultry litter pyrolyzed at 330°C, including N-heteroc organic compounds in the light and heavy fractions of poultry
litter pyrolyzed at 330° C, including N-heterocyclics, substituted furans, phenol and substituted phenols, benzenes and substituted benzenes, as well as aliphatic C chains. It is likely that some of these compounds remain trapped in the biochar pore structure, but few of these compounds appear to be immediately water soluble.

Ammonium‐nitrogen in the biochar leachate was also found in very low concentrations (table 4). No $NO₃-N$ was detected in any of the leachates. The NH_4-N concentrations were highest in the leachate from the PL biochar. Fresh poultry litter typically contains about 2.8 g NH₄-N kg⁻¹ (University of Georgia Agricultural and Environmental

Services Laboratory, unpublished data). Small amounts of this NH_4-N may remain trapped, or microbes may have mineralized nitrogen‐containing organic compounds in the biochar. Das et al. (2008) found that the liquid products obtained from poultry litter pyrolysis enhanced microbial growth in well water and concluded that N-heterocyclic compounds derived from proteins were responsible for that increase. A small fraction of these compounds may be present in the biochars.

INFLUENCE OF PYROLYSIS TEMPERATURE

The average conversion ratio (biochar weight / feedstock weight) was 33.2%. Biochar yield decreased with increased pyrolysis temperature, and except for N, nutrient weight) was 33.2%. Biochar yield decreased with increased
pyrolysis temperature, and except for N, nutrient
concentrations were higher in the biochar produced at 500° C (tables 2a andb). Due to the wide range of initial nutrient concentrations in the feedstock, there were significant interactions between feedstock and temperature for total N, P, Mg, Mn, Cu, Fe, Zn, and Al $(p < 0.05)$.

As noted earlier, N was conserved in the biochar (fig. 1b). After forest fires, on average, only 3% of the N in the biomass is found in ash, which contains black carbon or biochar (Giardina et al., 2000). Almendros et al. (2003) found C and N enrichment in charred residues during thermal transformation of peat organic matter. Nitrogen was incorporated into structures resistant to heating at moderate thermal oxidation by aromatization and formation of heterocyclic N (Almendros et al., 2003). Studies of wildfire effects on biomass composition indicate that N begins to the
terocyclic N (Almendros et al., 2003). Studies of wildfire
effects on biomass composition indicate that N begins to
volatilize at 200 $^{\circ}$ C, and above 500 $^{\circ}$ C half of the N in organic matter is lost to the atmosphere. Our study indicated that a relatively high proportion of the feedstock N was conserved at low pyrolysis temperatures, and as expected more N was retained in the biochar at 400°C compared to 500°C (fig. 1b).

Knicker et al. (2005) has shown that fire and carbonization can increase the N content of SOC, but the alterations in chemical structure have long‐term consequences for N availability (Knicker and Skjemstad, 2000). Field trials of PN and PC biochar as a soil amendment with corn (*Zea mays*) indicate that PN biochar N is not plant available (Gaskin et al., 2007). However, Tagoe et al. (2008) studied N recovery of 15N‐labeled chicken manure and did not find differences in N availability between carbonized and dried chicken manure.

Figure 2. Representative relationship of the ratio of nutrient (K) in the biochar to feedstock and conversion efficiency for pine chip, peanut hull, and poultry litter biochars. Solid circles represent means, and bars indicate standard errors.

The total concentration of other elements (P, K, Ca, and Mg) significantly increased with increasing volatization losses of C, H, O, and N (tables 2a and 2b). Potassium is representative of the nutrient concentration seen (fig. 2). Potassium and P vaporize at temperatures above 760°C, S
Potassium and P vaporize at temperatures above 760°C, S representative of the nutrient concentration seen (fig. 2).
Potassium and P vaporize at temperatures above 760° C, S
and Na need temperatures above 800° C, and Mg and Ca are Potassium and P vaporize at temperatures above 760° C, S
and Na need temperatures above 800° C, and Mg and Ca are
lost only at temperatures above 1107° C and 1240° C, respectively (Lide, 2004, reviewed by Knicker, 2007). There was a significant interaction between temperature and feedstock for Mehlich I extractable concentrations of these elements ($p = 0.05$). At the low nutrient concentrations seen in the PC biochar, temperature appeared to have little effect. In the PN and PL biochars, Mehlich I extractable nutrients tended to decrease with increasing temperature. Mehlich I extractable Al and Fe were significantly increased in the tended to decrease with interest with the strategies with interest and Fe w

The CEC of biochar produced at 500°C was significantly less than that produced at 400° C (table 2a, p < 0.01). There was a significant interaction between feedstock and temperature. In general, the literature indicates the loss of surface functional groups with the increase in pyrolysis temperature. Guo and Rockstraw (2007b) showed that the number of acidic functional groups decreased with increasing temperature. The highest decrease occurred between 300°C and 400°C, and the loss of these acidic
between 300°C and 400°C, and the loss of these acidic increasing temperature. The highest decrease occurred
between 300°C and 400°C, and the loss of these acidic
groups slowed after 400°C. This process may have contributed to the lower CEC seen at higher temperatures. Iyobe et al. (2004) indicated that lignin and cellulose undergo groups slowed after 400°C. This process may have
contributed to the lower CEC seen at higher temperatures.
Iyobe et al. (2004) indicated that lignin and cellulose undergo
thermolysis at 400°C to 500°C, which creates acidi functional groups such as carboxyls and phenolic hydroxyls. Chun et al. (2004) found decreasing acidity and increasing basicity with increasing pyrolysis temperature.

Temperature influenced DC (table 4). The higher temperature reduced the concentration of organic C but increased inorganic C significantly.

INFLUENCE OF STEAM ACTIVATION

Steam activation had little effect on the studied parameters (tables 2a, 2b, 3a, 3b, and 4). Production technology is known to influence physical parameters, and steam can improve the yield and surface characteristics at elevated pressures and temperatures (Antal and Grønli, 2003). At the relatively low pyrolysis temperatures used in this study, we only found significantly higher C and Mehlich I extractable B concentrations in steam‐activated biochar $(p < 0.05)$.

CONCLUSIONS

Pyrolytic biochar has the potential to be used in agricultural production to sequester carbon and serve as a fertilizer. Although pyrolysis conditions are known to affect the chemical and physical characteristics of biochar, at the relatively low pyrolysis temperatures used in this study, feedstock characteristics had the greatest influence on key agricultural characteristics. Carbon concentrations in the biochars decreased with increasing mineral content of the feedstock. Little DC was leachable from the fresh biochar. A high proportion of the feedstock N was conserved in the biochar; however, the N may not be plant available. Nutrients such as P, K, and Ca are extractable with a weak double acid extractant and may be plant available.

The higher pyrolysis temperature increased nutrient concentrations, except for N, but decreased CEC. Recent literature has shown that natural long‐term oxidation of biochar in the soil increases the amount of negative charges on the biochar surface (Cheng et al., 2008). Development and optimization of pyrolysis and post‐production treatments to increase CEC or available nutrients is important in order to increase the immediate benefits of biochar applications in agriculture.

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